MASENO UNIVERSITY S.G. S. LIBRARY

Mathematics of Pesticide Adsorption in a Porous Medium:

Convective—Dispersive Transport with steady state water flow

In

Two Dimensions

By

ADAMS SETH HARRISON WETOYI

A Thesis submitted in partial fulfillment of the requirement for the

Degree of Master of Science in Applied Mathematics

Department of Mathematics And Applied statistics

Maseno University

Kenya

February 2007

Abstract

The transport of solutes through porous media where chemicals undergo adsorption or change process on the surface of the porous materials has been a subject of research over years. Usage of pesticides has resulted in production of diverse quantity and quality for the market and disposal of excess material has also become an acute problem. The concept of adsorption is essential in determining the movement pattern of pesticides in soil in order to asses the effect of migrating chemical, from their disposal sites, on the quality of ground water. In the study of movement of pesticides in the soil, the mathematical models so far developed only consider axial movement. The contribution of radial movement to the overall location of solutes in the porous media seems to have been disregarded by researchers in this field. The objective of this study is to close this gap by developing a mathematical model to determine the combine radial and axial movement of pesticides due to Convective – Dispersive transport of pesticides with steady – state water flow in a porous media.

The methodology will involve determining the comprehensive dispersion equation accounting for both axial and radial movement of solutes in the porous media and finding the solution of the governing equation using finite difference methods. The solution of this equation will be applied to the data from experiments carried out on adsorption and movement of selected pesticides at high concentration by soil department, University of Florida, Gainesville U.S.A. We will confine our study to single – Region Flow and Transport.

CHAPTER 1

1.0 INTRODUCTION

1.1 Background to the problem

The effect of pesticides and other related contaminants of ground water have created concern to users and specialists. Contamination of ground water is caused by the transport of solutes through porous media. This process involves adsorption or change process on the surfaces of porous material.

Given the high cost of farming and the danger posed by pests on agricultural products, usage of pesticide is part and parcel of effective farming practice. This usage of pesticides has resulted in production of diverse quality for the market and the disposal of excess material is an acute problem.

Possible procedures of disposing pesticides include: incineration, encapsulation, isolation in ground caves and mines, chemical stabilization, land spread and land filling. Of all these methods, disposal by landfills and land spreading appear to be more economical [12]. The disheartening fact is that these hazardous chemical disposals do not provide a guarantee that the disposed material will not migrate from the disposal site to the ground water. We therefore require versatile comprehension of various processes that influence the persistence, retention and leaching of these hazardous chemicals in the soil so that the right chemicals with less migration potential can be recommended for use without jeopardizing life. Organic pesticides are preferred to inorganic pesticides because most organics have high degradation potential thus leading to their being considered less dangerous as compared to inorganic pesticides, which can stay in the soil environment for a very long time.

Our focus will be based on adsorption of pesticides and other solutes while moving through a porous medium.

The concept of adsorption is essential in determining the movement patterns of pesticides in the soil because it helps in assessment of the effect of migrating chemicals on the quality of the ground water environment.

Purity of ground water has become of increased concern given the diminishing availability of safe usable water.

Ground water pollution may be defined as artificially induced degradation of natural ground water quality.

Pollution can impair the use of water and create hazard to public health through toxicity or spread of disease [6]

Ground water pollution may go undetected for years, while remediation is difficult and costly. Any attempt to evaluate ground water pollution requires an understanding of particular aquifer system, its discharge and pollution pathways [14]

1.2 Statement of the problem

Given that no known study of convective-dispersive transport with steady state water flow gives us two dimensions equation for assessing pesticides movement in a porous media, we have to find a mathematical formula to help in determining this movement.

1.3 Objective of study

All known equations that describe movement of solutes through porous media, only takes into account the movement in axial direction. Figure 1 below shows that, when we use the existing equations we will conclude that there is no pollution caused by dumping of pesticides in a pit located some distance from the borehole site. In this situation the dumping site has an impervious rock under it (i.e. acquiclude layer). Use of equation (1) due to Van Genuchten [16] below in the prediction of pesticide effect on the borehole will not be possible.

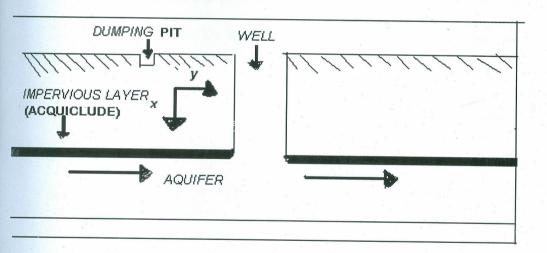


Figure 1 X-section of the rock strata

$$\frac{\partial C}{\partial t} = D \frac{\partial^2 C}{\partial x^2} - V \frac{\partial C}{\partial x} - \frac{\rho}{\theta} \frac{\partial S}{\partial t}$$
 (1.0)

in which C is the concentration of solute, D is the dispersivity constant, V is the pore water velocity, ρ is the bulky density of the soil, S is amount sorbed by solid phase, θ is volumetric water content and t is time.

Seemingly it is possible to assume that the pesticides in dumping pit have no effect on the borehole. However, this is untrue. It is therefore necessary to develop an appropriate model to fill this gap.

1.4 Significance of the study

Most studies carried out on adsorption of chemical solutes in soils or porous media only take into account axial movement disregarding horizontal movement of these solutes. In this study we have taken care of both horizontal and axial movement of chemical solute in the subsurface environment thus coming up with comprehensive model describing the movement of these solutes.

The study is essential to:

- i. The users so that they can determine the appropriate dumping site for given pesticides safeguarding the borehole or well.
- ii. The manufacturers so that they can manufacture pesticides with low migration potential from the dumping site.
- iii. The researchers so that they can be able to further carry out more studies so that they can provide sound advise to the user and the manufacturer and expand the knowledge base.

CHAPTER 2

2.0 LITERATURE REVIEW.

Taylor [5], in 1953 in his dispersion paper considers the diffusion of solute in a section through which poiseneille flow passes (i.e. laminar flow). If the mean velocity is U and the tube has radius α , the velocity V is described by

$$V = 2U(1 - \frac{r^2}{a^2}) , (2.0)$$

where r is the radial coordinate, a is the diameter of the tube.

Also the concentration C satisfies the equation

$$\frac{\partial C}{\partial t} + 2U(1 - \frac{r^2}{a^2})\frac{\partial C}{\partial t} = D\frac{\partial^2 C}{\partial r^2} + \frac{1}{r}\frac{\partial C}{\partial t} + \frac{\partial^2 C}{\partial x^2} \quad , \tag{2.1}$$

where *x* is the longitudinal measurement along the tube and t is time.

Taylor [5], in 1956 showed that when Peclet number which is denoted P_e and defined by $P_e = \frac{aU}{D}$ is large (Peclet number is one of parameters used in assessing dispersivity), the effect of diffusion term in equation (2.1) is to dispersed the mean solute concentration diffusively D^* about the position of its centre of mass Ut = x with dispersion coefficient

$$D^* = \frac{a^2 U}{48D} {.} {(2.2)}$$

Aris [5] later improved the above equation to

$$D_T = \frac{a^2 U}{48 D_L} + D_L \quad , \tag{2.3}$$

where D_T is the total dispersion coefficient, $D = D^* + D_L$, D^* is the molecular diffusion constant and D_L is the longitudinal dispersion constant which is valid for $\frac{U_a}{D_L} \ge 1$. The dispersion mechanism is due to the radial variation of velocity profile, which disperses the solute even if the diffusion is small.

The turtuosity (lack of straight forwardness) of the flow paths and the possibility of the adsorption on the solids surface cause D* to be less than D and the ratios of the $\frac{D^*}{D}$ between 0.001 and 0.5 are commonly observed [5]. In porous media, remixing at pore junction causes dependence of D_L on flow velocity to the less quadratic and relation of the form

$$D_{L} = \alpha_{L} V^{m} \quad , \tag{2.4}$$

where 1 < m < 1.2 and α_L is a constant of Dispersivity.

Mixing at junctions also causes transverse dispersion to occur with the dispersion coefficient D_T (In transverse direction) that measures to be less than D_L by a factor of order 10, when $P_e >> 1$ [20]

the Peclet number shows little variation with the increase in Reynolds number in liquids and is of order of unity [20]

The Reynolds number is given by

$$Re = \frac{U_C d\rho}{\mu} \quad , \tag{2.5}$$

where U_c is critical velocity, d is diameter of particle, ρ is density of the fluid and μ is the viscosity of the fluid

(Reynolds number is another parameter used in assessment of fluid flow).

Darcy [14], in 1856 described flow of water through homogenous layers in saturated soil as

$$q = \frac{Q}{At} = -K\frac{\partial H}{\partial z} , \qquad (2.6)$$

where q is the flux density (LT⁻¹), Q is the volume of water (L³) passing through a cross section area A (L²) per unit time, H(L) is the hydraulic head and is the sum of the gravitational head z(L) and pressure head h(L) and K is the hydraulic conductivity (LT⁻¹), which represents the ability of the soil to conduct water and is considered to be constant under saturated conditions. Equation (2.6) applies to non-steady state, unsaturated flow as in [15]

$$q = -k(h)\frac{\partial H}{\partial z},\tag{2.7}$$



where K is a function of h (L), the pressure head.

Van Genuchten et al [26] developed a partial differential equation generally assumed to describe the movement of pesticides and other adsorbed solutes through soils under steady state water flow condition as

$$\frac{\partial C}{\partial t} = D \frac{\partial^2 C}{\partial x^2} - V \frac{\partial C}{\partial x} - \frac{\rho}{\theta} \frac{\partial S}{\partial t}.$$
 (2.8)

Later Van Genuchten and Alves [27] came up with an equation for one dimensional, miscible displacement of an absorptive and degradable chemical species

$$\frac{\partial}{\partial x} \left[D \frac{\partial C}{\partial x} \right] - V \frac{\partial C}{\partial x} - R \frac{\partial C}{\partial t} = \mu C - \gamma , \qquad (2.9)$$

where R is the retardation factor, μ is the decay constant and γ is zero order production rate.

Basically the equation applies to degradable organic pesticides and not un-degradable inorganic pesticides.

Langmuir [4], in 1915 came up with a model, which is also called ideal localized monolayer model, which was based on the assumptions below:

- i. The molecules are absorbed on definite site
- ii. Each site can accommodate only one molecule
- iii. The area of each site is fixed quantity determine solely by the geometry of the surface.
- iv. The adsorption energy is the same at all site
- v. The adsorbed molecules cannot migrate across the surface or interact with neighbouring molecules

The equation gave

$$S = \frac{X_m b C_e}{1 + b C_e} \quad , \tag{2.10}$$

S= $\frac{x}{m}$ is the amount of solute adsorbed x per unit weight of adsorbent m, C_e is equilibrium concentration of the solute adsorbed per unit weight of absorbent required for monolayer coverage of the surface, also called Monolayer capacity, b is a constant related to the heat of adsorption Q $\left[b\alpha ex\rho(-\frac{\Delta H}{RT})\right]$ [4].

Freundlich [20], in 1926 came up with adsorption equation, which is widely used as a mathematical description of adsorption in aqueous system

$$S = S_e + S_k + S_{ir} {,} {(2.11)}$$

where S_e is adsorption governed by instantaneous equilibrium reactions, S_k is adsorption governed by hysteretic kinetic reaction, S_{ir} is adsorption subject to irreversible retention.

i. The Langmuir adsorption equation.

The instantaneous equilibrium reaction between the amount of chemical in solution and that sorbed by the solid phase is generally represent by one of the following three adsorption isotherms

The simplest chemical reaction model is the linear adsorption equation

$$S_e = K_d C, (2.12)$$

where K_d is referred to as the distribution coefficient (slope of the adsorption isotherm)

ii. Equation (2.13) is a special case of the Freundlich equation

$$Se = K_d C^N , (2.13)$$

where N is a fitting parameter.

iii. The Lagmuir adsorption equation,

$$S_e = \frac{aC}{1+bc} , \qquad (2.14)$$

where a and b are curve fitting parameters.

BASIC CONCEPTS

3.1 Adsorption

surface tension. In any solid or liquid, atoms at the surface are subject to unbalanced forces of attraction normal to the surface. These forces are merely an extension of acting forces within the body of the material. A molecule in the centre of a liquid drop is attracted from all sides. While at the surface the attractive acting between adjacent molecules result in a net attraction in the bulky phase in the direction normal to the surface. Because of the unbalanced attraction at the surface there is a tendency of these molecules to be pulled from the surface into interior and for the surface to shrink to the smallest area that can enclose the liquid [18]. Adsorption is promoted by charged clay mineral, hydrous oxide coating on surfaces and organic matter functional groups with variable charges. Additionally, some solutes may co-precipitate, volatilize or degrade [32]

Adsorption is a natural process by which fluids are attracted and then held at the surface of solid caused by

3.1.1 Physical adsorption

The adsorption process may be classified as physical or chemical.

Physical adsorption on solids is attributed to the forces of interaction between the solid surface and the adsorbent molecules that are similar to Van der Waals forces between the molecules. As these non-bonded molecules approach each other more closely, these attractive forces vanish and strong repulsive forces emerges, which cause a sharp rise in energy content of molecules making it less stable. These forces that include the electron and the nuclei of a system are electrostatic in origin and are termed as dispersive forces. The dispersive forces exist in all types of matter and always act as attractive force between an adjacent atom and molecule no matter how dissimilar. They are always present despite the nature of the other adsorbate-adsorbent potential [1]

The nature of the dispersive forces was first recognized in the 1930 by London Van der Waals [3]. Using Quantum mechanical calculation it was postulated that electron motion in an atom or molecule would lead to dipole oscillating dipole movement. At any instance, the lack of symmetry of the electron distribution about the

nuclei imparts a transient dipole moment to an atom or molecule that would average to zero over a long time interval. When in close proximity to a solid surface each instantaneous dipole of an approaching molecule induce an approximately oriented dipole moment in a surface molecule. These forces are known as dispersion forces because of this relationship. The dipole dispersion interaction energy, E, can be determined by

$$E = -\frac{C}{r^2} \,, \tag{3.0}$$

where C is a constant and r is the distance of separation between the interacting molecules.

In additional to dipole—dipole interactions other possible dispersion contributing to physical adsorption includes dipole—Quadra pole and Quadra pole interaction. If these are included, the total dispersion energy is given by

$$E = -\frac{C}{r^6} - \frac{C_1}{r^8} - \frac{C_2}{r^{10}} \tag{3.1}$$

where C_1 is a constant of dipole – Quadra pole interaction, C_2 is a constant of Quadra pole interactions. The overall interactions is expressed as

$$E = -\frac{C}{r^6} + \frac{B}{r^{12}} \tag{3.2}$$

where B is a repulsive interaction constant.

3.1.2 Chemical adsorption

This is characterized mainly by large interaction potentials that lead to high heats of adsorption approaching the value of chemical bonds. It involves transfer of electrons and formation of true chemical bonding between the adsorbate and the solid surface [4]. The high temperature associated with chemisorption is usually associated with activation energy released in chemical bonding. In chemisorption the adsorbed molecules remain in situ. (i.e. site specific)

3.1.3 Factors affecting adsorption

Nature of the adsorbent:

i. Surface area and pore structure. – Adsorption of a solid adsorbent is directly proportional to the specific surface area and pore size distribution.

The more the surface area is accessible to the sorbate the high the adsorption.

- ii. Particle size The lower the particle diameter the higher the adsorption. But for very highly porous adsorbents most surface area reside in internal pore structure; therefore the adsorptive capacity is independent of the particle size. [35]
- iii. Chemistry of the surface The presence of specific functional groups on the surface of adsorbent affects the adsorption process.

Nature of the adsorbate:

- i. In general, a higher solubility indicate a strong solute solvent interaction or affinity the extend of adsorption is expected to be low due to the necessity of breaking the solute solvent bond before adsorption occur.
- ii. The molecular weight and size of the adsorbate molecule also affect the adsorptive capacity [30]. The adsorptive capacity increases with an increase in mole weight.

Effect of temperature

i. Since the process of adsorption is spontaneous, it is accompanied by decrease in entropy due to loss of degrees of freedom of the solute in passing from the dissolved state to the adsorbate state [34] The adsorption process is always exothermic therefore an increase in temperature will result in a reduction of equilibrium adsorptive capacity and otherwise for lower temperature.

3.2 Adsorption from liquids

The presentation of the amount of solute adsorbate per unit adsorbent as a function of the equilibrium concentration in bulky solution at a constant temperature is termed as the adsorption isotherm.

One of the most popular adsorption isotherm equations that is used for liquids was described as

$$S = KC_e^{\ N}, \tag{3.3}$$

where $S = \frac{x}{m}$, is adsorbed solid and C_e is the solute equilibrium constant.

The Freundlich equation is an empirical expression that encompasses the heterogeneity of the substance and the

exponential distribution of sites and their energies.

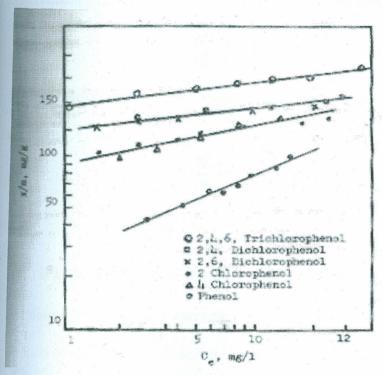


Fig 2: Logarithmic form of Freundlich adsorption isotherms for phenolic compound on activated carbon For linearization of the data in equation (3.3) can be expressed as

$$\log S = \log K + N \log C$$

and plotting $\log S$ against C enables one to find N and K(see figure 2).

Steep slopes indicate high adsorptive capacity at equilibrium concentration that rapidly diminishes at lower concentration. Relatively flat slopes i.e. $N \ll 1$, indicate that the adsorptive capacity is only slightly reduced at lower equilibrium concentration.

3.3 Solution of one dimension equation

The partial differential equation assumed to describe the movement of pesticides and other solutes through soils under steady water flow condition is equation (2.8).

We shall solve the equation using numerical method in order to compare our subsequent results with twodimensional equation which we intend to derive and solve. In our study we shall use numerical methods to solve equation (2.8). The limitation of the equation is that it is in one dimension. We are going to develop a model for solving this equation numerically.

From equation (3.3), $S = KC^N$, since S is a function of another variable C

$$\frac{\partial S}{\partial t} = \frac{\partial S}{\partial C} \frac{\partial C}{\partial t} = NKC^{N-1} \frac{\partial C}{\partial t}.$$
(3.5)

When the adsorption isotherm obeys the Freundlich equation, the convective-dispersive solute transport model equation (2.8) reduces to:

$$R(C)\frac{\partial C}{\partial t} = D\frac{\partial^2 C}{\partial x^2} - V\frac{\partial C}{\partial x},$$
(3.6)

where R (C) =
$$(1 + \frac{\rho N K_d C^{N-1}}{\theta})$$
 (3.7)

R (C) is a retardation term and index of pesticide mobility, K_d is the Freudlich [26] adsorption constant. The retardation term R (C) is a quantitative index of pesticides' mobility in that its value is equal to the ratio of the position of the adsorbed value fronts. The value of the adsorption coefficient K in the equation (3.5) for non-adsorbed solutes is equal to zero hence R (C) = 1. For adsorbed solutes, R (C) is greater than unity since the value of K is larger than zero. The larger values of R (C) indicate reduced pesticides mobility in soils. It may be noted from equation (3.7) that for the case of non linear adsorption isotherms (N<1), the retardation term varies inversely with solution concentration C. For nonlinear isotherm (N=1), R (C) is independent of pesticide solution concentration. Thus the shape of the equilibrium adsorption isotherm directly influences the mobility of pesticide and other adsorbed solutes through the soil.

Consider equation (3.6), when rearranged

$$D\frac{\partial^2 C}{\partial x^2} - V\frac{\partial C}{\partial x} - R(C)\frac{\partial C}{\partial t} = 0$$
(3.8)

When a physical system depends on more than one variable, a general description of its behavior often leads to a equation containing partial differentials. Equation (3.6) is our partial differential equation and we solve it using finite difference method. The calculus of finite differences will enable solve this differential equation numerically

by calculating the values of the function at discrete (finite) points.

The finite difference method is ideal for solving non linear equations. We replace the differential with its finite difference equivalent. We shall establish grids based on dimensions we are to consider. We use the (i, j) notation that is used to designate the pivot point for two-dimensional space (x, y) direction and (i, j) being the counters in the (x, y) directions. The partial derivative of C with respect to x implies that t is kept constant and vice versa. Where n is an arbitrary Rx(0,T) boundary condition, where $R=a \le x \le b$ and $0 \le t \le T$.

The initial condition is that the concentration of pesticide at all positions in the soil at time zero is constant and equal to C_i . That is $C(x,0) = C_i$ for x > 0.

Boundary conditions: two conditions are necessary:

i. In the first case the concentration of the pesticides at the position x = 0 is specified for a period of time, the concentration at the surface is zero. That is

$$C(0,t) = C_0 \text{ for } 0 < t \le t_0$$

$$C(0,t) = 0 \text{ for } t > t_0$$

ii. In the second case, the concentration of the pesticides in the solution entering the soil system at position x = 0 is specified for a period time. Following that time, the concentration at the surface is zero. Mathematically this is written as,

$$-D\frac{dC}{dx} + VC|_{x=0} = \begin{vmatrix} VC_{0}, & \text{for, } 0 < t \le t_{0}. \\ 0, & \text{for, } t > 0. \end{vmatrix}$$

Assumptions

- i. The pore water velocity is constant in time and space. This condition can be met for a uniform soil if the flux density of water velocity and volumetric water content are constant for all positions all the times.
- ii. The spread of solute is dominated by hydraulic dispersion rather than diffusion.
- iii. The hydrodynamic dispersion can be approximated as the product of the dispersivity and pore water velocity.
- iv. The adsorption process is instantaneous and reversible and the adsorption isotherm can be described by

the model i.e the concentration of pesticide absorbed on the soil solids is proportional to the concentration in the solution,[14]

When we approximate the differential $\frac{\partial C}{\partial x}$ using central differences, we find $h = \Delta x$

We now consider equation (3.8). The coefficient R(C), is a function of a dependent variable C, therefore the equation is quasilinear. For a properly posed initial value problem and its finite difference equation to satisfy the consistency condition, stability is necessary and sufficient condition for convergence.

Using forward difference,

taking $k = \Delta t$,

$$\frac{\partial C}{\partial x} = \frac{1}{h} (C_{i+1}^n - C_i^n), \tag{3.9}$$

$$\frac{\partial C}{\partial t} = \frac{1}{k} (C_i^{n+1} - C_i^n), \tag{3.10}$$

$$\frac{R(C_i^n)}{k} \left\{ C_i^{n+1} - C_i^n \right\} = \frac{D}{h^2} \left\{ (C_{i+1}^n - 2C_i^n + C_{i-1}^n) \right\} - \frac{V}{h} (C_{i+1}^n - C_i^n). \tag{3.11}$$

From equation (3.8)

$$C_{i}^{n+1} = \left(\frac{kD}{R(C_{i}^{n})h^{2}} - \frac{kV}{R(C_{i}^{n})h}\right)C_{i+1}^{n} + \left(1 + \frac{kV}{R(C_{i}^{n})h} - \frac{2kD}{R(C_{i}^{n})h^{2}}\right)C_{i}^{n} + \frac{kD}{R(C_{i}^{n})h^{2}}C_{i-1}^{n}$$
(3.12)

Explicit methods similar to the one used in solving our equation above are computationally easier to solve, however, since our equation will become more complex when we introduce the second dimension to the above equation, we shall use implicit techniques because they are unconditionally stable while explicit techniques are conditionally stable. Now consider equation (3.8). We utilize the grid of figure 3 at half point in the n-direction (i, n + 1/2). Instead of expressing in terms of forward difference around (i, n) as it was done in the explicit computation, we express $\frac{\partial C}{\partial t}$, in terms of forward differences around the half point,

$$\frac{\partial C}{\partial t}\Big|_{i,n+\frac{1}{2}} = \frac{1}{k} \left(C_i^{n+1} - C_i^n\right) \tag{3.13}$$

The first and second order partial derivatives are expressed at the half point as a weighted average of the central differences at points (i, n+1) and (i, n) as under:

$$\frac{\partial C}{\partial x}\Big|_{i,n+\frac{1}{2}} = p\partial C_{i}^{n+1} + (1-p)\partial C_{i}^{n},$$

$$= \frac{1}{2h} \left[pC_{i+1}^{n+1} - pC_{i-1}^{n+1} + C_{i+1}^{n} - C_{i-1}^{n} - pC_{i+1}^{n} + pC_{i-1}^{n} \right].$$

$$\frac{\partial^{2}C}{\partial x^{2}}\Big|_{i,n+\frac{1}{2}} = p\partial^{2}C_{i}^{n+1} + (1-p)\partial^{2}C_{i}^{n},$$

$$= p\left[\frac{1}{h^{2}} \left(C_{i+1}^{n+1} - 2C_{i}^{n+1} + C_{i-1}^{n+1} \right) \right] + \left(1-p \right) \left[\frac{1}{h^{2}} \left(C_{i+1}^{n} - C_{i}^{n} + C_{i-1}^{n} \right) \right].$$

$$\frac{(i-1, n+1)}{(i,n+\frac{1}{2})} \frac{(i,n+1)}{(i,n+2)} \frac{(i+1, n+1)}{(i,n+2)}$$

(i+1,n)

Figure 3: Display of nodal points.

Where p is in the range of, $0 \le p \le 1$.

Inserting the above expressions in equation (3.8) yields:

(i-1,n)

$$\frac{R(C_{i}^{n})}{k} \left(C_{i}^{n+1} - C_{i}^{n}\right) = \frac{D}{h^{2}} \left[p \left(C_{i+1}^{n+1} - 2C_{i}^{n+1} + C_{i-1}^{n+1}\right) + (1-p) \left(C_{i+1}^{n} - 2C_{i}^{n} + C_{i-1}^{n}\right) \right] - \frac{V}{2h} \left[p \left(C_{i+1}^{n+1} - C_{i-1}^{n+1} - C_{i+1}^{n} + C_{i-1}^{n}\right) + \left(C_{i+1}^{n} - C_{i-1}^{n}\right) \right].$$

(i,n)

Using Schmidt method [12] i.e. p=0 we find:

$$C_{i}^{n+1} = \frac{k}{R(C_{i}^{n})h} \left(\frac{D}{h} - \frac{V}{2}\right) C_{i+1}^{n} + \left(1 + \frac{2kD}{R(C_{i}^{n})h^{2}}\right) C_{i}^{n} + \frac{k}{R(C_{i}^{n})h} \left(\frac{D}{h} + \frac{V}{2}\right) C_{i-1}^{n},$$

which is unconditionally stable. When p=1, the system is fully implicit.

CHAPTER 4

4.0 DERIVATION OF CONVECTIVE-DISPERSIVE SOLUTE TRANSPORT EQUATION WITH STEADY STATE WATER FLOW CONDITION

Average pore water velocity $V(LT^{-1}) = \frac{q}{\theta}$,

i.e. $q = -K \frac{\partial H}{\partial z}$, is the flux density, (from equation 2.6), $\theta = \frac{V_W}{V_S}$, in which V_W is volume of water in the porous media and V_S is volume of solids is used instead.

In this study we are going to use the concept of dispersion through a cylindrically packed soil vessel to derive our equation.

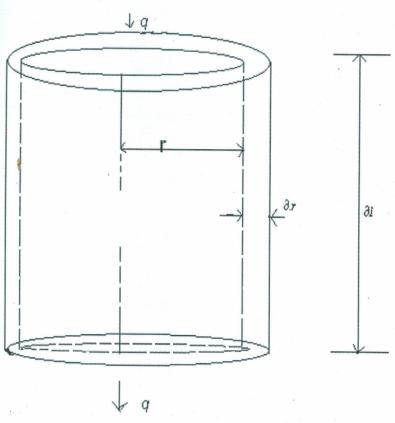


Figure 4: Derivation of the equation.

At very low flow rate, the dispersion is different in longitudinal and radial directions. The Dispersion coefficients denoted by D_L for longitudinal and D_R for radial

 $D(\theta, V) = D_{diff} + D_{dis}, (4.0)$

where Ddiff (L^2T^{-1}) is molecular diffusion coefficient, Ddis (L^2T^{-1}) is the hydrodynamic dispersion and is the mixing or spreading of the solute during transport due to differences in velocities within a pore and between pores. The volumetric water content denoted by θ which we can assume to be the voidage for saturated soils. The element height is denoted by $\partial \ell$. Inner radius is r and outer radius is $r + \partial r$, C is the concentration of the material to be dispersed and is a function of axial position l, radial position r, time t and dispersion coefficients D_R and D_L radial and axial respectively.

The rate of entry of reference material due to flow in axial direction is $q(2\pi r\partial rC)$. The corresponding efflux rate

is

$$q\left(2 \pi r \partial r\right) \left(C + \frac{\partial C}{\partial l} \partial l\right). \tag{4.1}$$

The net accumulation rate in element due to flow in axial direction is:

$$-q\left(2\,\pi r\partial r\right)\left(\frac{\partial C}{\partial l}\,\partial l\right).\tag{4.2}$$

Rate of diffusion in axial direction across inlet boundary is:

$$-\left(2\pi r\partial r\ \theta\right)\left(D_{L}\frac{\partial C}{\partial l}\right). \tag{4.3}$$

The corresponding rate at outlet boundary is:

$$(2\pi r\partial r \ \theta) \ D_L \left(\frac{\partial C}{\partial l} + \frac{\partial^2 C}{\partial l^2} \partial l \right). \tag{4.4}$$

The net accumulation rate due to diffusion from boundaries in axial direction is:

$$(2\pi r \partial r \ \theta) D_L \frac{\partial^2 C}{\partial l^2} \partial l \ . \tag{4.5}$$

Diffusion in radial direction at r is:

$$-(2\pi r\partial r \ \theta) \partial lD_R \frac{\partial C}{\partial r}$$
 (4.6)

The corresponding rate at radius $r + \partial r$ is

$$\left[2\pi(r+\partial r)\theta\right]\partial l \ D_{R}\left[\frac{\partial C}{\partial r} + \frac{\partial^{2}C}{\partial r^{2}}\partial r\right]. \tag{4.3}$$

The net accumulation rate due to diffusion from boundaries is:

$$-\left[2\pi r\partial r\theta\right]\partial l \quad D_{R}\frac{\partial C}{\partial r} + \left[2\pi(r+\partial r)\partial l(\theta)\right]D_{R}\left(\frac{\partial C}{\partial r} + \frac{\partial^{2} C}{\partial r^{2}}\partial r^{2}\right). \tag{4.8}$$

If we ignore the last term, it becomes:

$$2\pi\theta D_R \partial l \left[\partial r \frac{\partial}{\partial r} \left(r \frac{\partial C}{\partial r} \right) \right]. \tag{4.9}$$

For a representative elementary volume of soil, the total amount of a given chemical species X (ML⁻³) is represented by the sum of the amount retained by the soil matrix and the amount present in the soil.

$$X = \rho_b S + \theta C \,, \tag{4.9}$$

where, ρ_{r} is the bulky density, and S is the amount of solute adsorbed,

$$\frac{\partial X}{\partial t} = \rho_b \frac{\partial S}{\partial t} + \theta \frac{\partial C}{\partial t} .$$

(4.10

Now the total accumulation rate is:

$$(2\pi r \partial r \partial l) \frac{\partial X}{\partial t}$$

$$= (2\pi r \partial r \partial l) \left(\rho_b \frac{\partial S}{\partial t} + \theta \frac{\partial C}{\partial t} \right). \tag{4.11}$$

Thus from equations (4.0) through to (4.11), we have:

$$\left(\rho_{b}\frac{\partial S}{\partial t} + \theta \frac{\partial C}{\partial t}\right) 2\pi r \partial r \partial l = -q(2\pi r \partial r)\frac{\partial C}{\partial l} \partial l + (2\pi r \partial r \theta)D_{L}\left(\frac{\partial^{2} C}{\partial l^{2}} \partial l\right) + 2\pi \partial l D_{R}\left[\partial r \frac{\partial C}{\partial r}\left(r \frac{\partial C}{\partial r}\right)\right], \quad (4.12)$$

and on dividing through by $(2\pi r \partial r)\partial l\theta$, we find

$$\left(\frac{\rho}{\theta}\frac{\partial S}{\partial t} + \frac{\partial C}{\partial t}\right) = D_L \frac{\partial^2 C}{\partial t^2} + \frac{1}{r}D_R \frac{\partial}{\partial r}\left(r\frac{\partial C}{\partial r}\right) - \frac{q}{\theta}\frac{\partial C}{\partial t}$$

(4.13

Taking l=x and r=y our equation comes to

$$\frac{\rho}{\theta} \frac{\partial S}{\partial t} + \frac{\partial C}{\partial t} = D_x \frac{\partial^2 C}{\partial x^2} + \frac{1}{y} D_y \frac{\partial}{\partial y} \left(y \frac{\partial C}{\partial y} \right) - \frac{q}{\theta} \frac{\partial C}{\partial x} .$$

(4.14

But $\frac{q}{\theta} = V_{\chi}$ (pore water velocity), therefore equation (4.8) comes to

$$\left(\frac{\rho}{\theta} \frac{\partial S}{\partial C} * \frac{\partial C}{\partial t} + \frac{\partial C}{\partial t}\right) = D_x \frac{\partial^2 C}{\partial x^2} + \frac{1}{y} D_y \frac{\partial}{\partial y} \left(y \frac{\partial C}{\partial y}\right) - V_x \frac{\partial C}{\partial x} \quad .$$
(4.15)

From the Freundlich equation, equation (3.4), we have

$$S = KC^{N}, \frac{\partial S}{\partial C} = KNC^{N-1}, \frac{\partial S}{\partial t} = \frac{\partial S}{\partial C} \frac{\partial C}{\partial t}, = KNC^{N-1} \frac{\partial C}{\partial t}. \tag{4.16}$$

Putting equation (4.16) in (4.15)

$$R(C)\frac{\partial C}{\partial t} = D_x \frac{\partial^2 C}{\partial x^2} - V_x \frac{\partial C}{\partial x} + \frac{1}{y} D_y \frac{\partial}{\partial y} \left(y \frac{\partial C}{\partial y} \right) ,$$

$$\text{where, } R(C) = \left(1 + \frac{\rho}{\theta} KNC^{N-1} \right).$$

$$(4.17)$$

Equation (4.17) is our model equation describing two-dimensional movement of solute in the soil or porous media.

CHAPTER 5

5.0 FINITE DIFFERENCES SCHEME TO THE TWO DIMENSION EQUATION

The partial differential we have derived describing the movement of pesticides and other solute through the soils under steady flow condition similar to the one dimensional of Van Genuchten (1974)'s one-dimension equation

is

$$R(C)\frac{\partial C}{\partial t} = D_x \frac{\partial^2 C}{\partial x^2} - V_x \frac{\partial C}{\partial x} + \frac{D_y}{y} \frac{\partial C}{\partial y} + D_y \frac{\partial^2 C}{\partial y^2}$$
(5.0)

The partial derivatives of C with respect to x, implies y and t are kept constant and vice versa.

i.e.

$$\frac{\partial C}{\partial x} \left| i, j, n = \frac{dC}{dx} \right|_{i, j, n} , \qquad (5.1)$$

$$\frac{\partial C}{\partial y} \bigg| i, j, n \equiv \frac{dC}{dy} \bigg|_{i,j,n},$$
 and,

$$\left. \frac{\partial C}{\partial t} \middle| i, j, n = \frac{dC}{dt} \middle|_{i,j,n} \right. \tag{5.2}$$

The initial condition; the concentration of pesticides at positions in the at time zero is constant,

and equal to C i, j, i.e. C(x,y,0) = Ci,j for x,y>0

Boundary conditions: two conditions are necessary.

- i. In the first case the concentration of pesticides at position x=0 and y=0 is specified for a period of time.
- ii. Following that time, the concentration at the surface is zero i.e.

$$C(0,0,t) = C_0$$
 for $0 < t \le t_0$ (5.3)

$$C(0,0,t) = 0$$
 for $t > t_0$

iii. In the second case, the concentration of pesticides in the solution entering the soil system at position x=0 and y=0 is specified for a period of time. Following that time the concentration at the surface is zero.

$$-\left(D_{x}\frac{\partial C}{\partial x}+D_{y}\frac{\partial C}{\partial y}\right)+VC\bigg|_{\substack{x=0\\y=0}}=\bigg|_{\substack{0 \text{ for } 0< t\leq t_{0}\\0 \text{ for } t>0}}^{VC_{0} \text{ for } 0< t\leq t_{0}}$$
(5.4)

Assumptions:

From our analysis in one dimension equation, we noted that when the coefficient is a function of dependent variables, the equation is quasilinear. We even used finite difference approximations to solve the one-dimensional equation describing the movement of pesticides in the porous media with respect to time.

We later used implicit method to find a numerical solution, which was dependably stable. In this analysis we are going to solve our two-dimensional partial differential equation using implicit methods to ensure our result are unconditionally stable. Starting with equation (5.0), i.e.

$$R(C)\frac{\partial C}{\partial t} = D_x \frac{\partial^2 C}{\partial x^2} - V_x \frac{\partial C}{\partial x} + \frac{D_y}{y} \frac{\partial C}{\partial y} + D_y \frac{\partial^2 C}{\partial y^2}$$

The problem with seeking solutions to the above equation is complicated by the presence of complex geometry. Solution by analytical means is complex, so we have to use numerical techniques to find the solution.

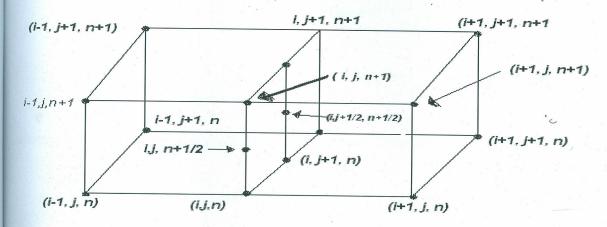


Figure 5: Two-dimensional nodal points

We utilize the grid of figure 5 in which the half a point in n-direction $(i, j, n + \frac{1}{2})$ is shown as K. Instead of expressing $\frac{\partial C}{\partial t}$ in terms of forward difference around (i, j, n), we express it in terms of central differences around the half point.

$$\left. \frac{\partial C}{\partial t} \right|_{i,j,n+\frac{1}{2}} = \frac{1}{k} \left(C_{i,j}^{n+1} - C_i^n \right) \tag{5.5}$$

These first and second order partial derivatives are expressed at the half point as weighted averages of the central differences at points (I, j, n+1) and (I, j, n)

Let $\Delta x=h$, $\Delta y=q$, $\Delta t=k$, then

$$\frac{\partial C}{\partial x} \bigg|_{i,j,n+\frac{1}{2}} = p\partial C_{i,j}^{n+1} + (1-p)\partial C_{i,j}^{n}
= \frac{1}{2h} \Big(pC_{i+1,j}^{n+1} - pC_{i-1,j}^{n+1} + C_{i+1,j}^{n} - pC_{i+1,j}^{n} + pC_{i-1,j}^{n} \Big).$$
(5.6)

$$\frac{\partial C}{\partial y}\bigg|_{i,j,n+\frac{1}{2}} = \frac{1}{2q} \Big(pC_{i,j+1}^{n+1} - pC_{i,j-1}^{n+1} + C_{i,j+1}^{n} - C_{i,j-1}^{n} - C_{i,j+1}^{n} + pC_{i,j-1}^{n} \Big).$$
(5.7)

$$\frac{\partial^{2} C}{\partial x^{2}} \bigg|_{i,j,n+\frac{1}{2}} = p \partial^{2} C_{i,j}^{n+1} + (1-p) \partial^{2} C_{i,j}^{n} ,$$

$$= \frac{1}{h^{2}} \left\{ \left[p(C_{i+1,j}^{n+1} - 2C_{i,j}^{n+1} + C_{i-1,j}^{n+1}) \right] + (1-p) \left[C_{i+1,j}^{n} - 2C_{i,j}^{n} + C_{i-1,j}^{n} \right] \right\}. \tag{5.8}$$

$$\frac{\partial^{2}C}{\partial y^{2}}\Big|i,j,n+\frac{1}{2} = p\partial^{2}C_{i,j}^{n+1} + (1-p)\partial^{2}C_{i,j}^{n},$$

$$= \frac{1}{q^{2}}\Big\{p\Big[C_{i,j+1}^{n+1} - 2C_{i,j}^{n+1} + C_{i,j-1}^{n+1}\Big] + (1-p)\Big[C_{i,j+1}^{n} - 2C_{i,j}^{n} + C_{i,j}^{n}\Big]\Big\}, \quad (5.9)$$

where p is in the range of $0 \le p \le 1$. Inserting the above finite difference expression in equation (5.0) we obtain:

$$\frac{R(C_{i,j}^{n})}{k} \left(C_{i,j}^{n+1} - C_{i,j}^{n} \right) = \frac{D_{x}}{h^{2}} \left[p \left(C_{i+1,j}^{n+1} - 2C_{i,j}^{n+1} + C_{i-1,j}^{n+1} \right) + (1-p) \left(C_{i+1,j}^{n} - 2C_{i,j}^{n} + C_{i-1,j}^{n} \right) \right]
- \frac{V_{x}}{2h} \left(p C_{i+1,j}^{n+1} - p C_{i-1,j}^{n+1} + C_{i+1,j}^{n} - C_{i-1,j}^{n} - p C_{i+1,j}^{n} + p C_{i-1,j}^{n} \right)
+ \frac{D_{y}}{y} \left[\frac{1}{2q} \left(p C_{i,j+1}^{n+1} - p C_{i,j-1}^{n+1} + C_{i,j+1}^{n} - C_{i,j-1}^{n} - p C_{i,j+1}^{n} + p C_{i,j-1}^{n+1} \right) \right]
+ \frac{D_{y}}{q^{2}} \left[p \left(C_{i,j+1}^{n+1} - 2C_{i,j}^{n+1} + C_{i,j-1}^{n+1} \right) + (1-p) \left(C_{i,j+1}^{n} - 2C_{i,j}^{n} + C_{i,j-1}^{n} \right) \right].$$
(5.10)

Again using Schmidt method i.e. p = 0 our equation reduces to

$$\begin{split} R\Big(C_{i,j}^{n}\Big)\Big(C_{i,j}^{n+1}-C_{i,j}^{n}\Big) &= \frac{D_{x}}{h^{2}}\Big(C_{i+1,j}^{n}-2C_{i,j}^{n}+C_{i-1,j}^{n}\Big) - \frac{V_{x}}{2h}\Big(C_{i+1,j}^{n}-C_{i-1,j}^{n}\Big) + \frac{D_{y}}{2yq}\Big(C_{i,j+1}^{n}-C_{i,j-1}^{n}\Big) \\ &\quad + \frac{D_{y}}{q^{2}}\Big(C_{i,j+1}^{n}-2C_{i,j}^{n}+C_{i,j-1}^{n}\Big), \end{split}$$

it eventually gives us

N and M are the largest value of i, j and n

When $\Delta x = \Delta y = h$ equation (5.11) becomes

$$C_{i,j}^{n+1} = \left[\frac{k}{R(C_{i,j}^n)h} \left(\frac{D_x}{h} - \frac{V_x}{2} \right) \right] C_{i+1,j}^n + \left[1 - \frac{2kD_x}{R(C_{i,j}^n)h^2} - \frac{2kD_y}{R(C_{i,j}^n)h^2} \right] C_{i,j}^n + \left[\frac{k}{R(C_{i,j}^n)h} \left(\frac{D_x}{h} + \frac{V_x}{2} \right) \right] C_{i-1,j}^n + \left[\frac{kD_y}{R(C_{i,j}^n)h} \left(\frac{1}{h} + \frac{1}{2y_i} \right) \right] C_{i,j+1}^n + \left[\frac{kD_y}{R(C_{i,j}^n)h} \left(\frac{1}{h} - \frac{1}{2y_i} \right) \right] C_{i,j-1}^n$$
(5.12)

Let:

$$\mathbf{A} = \left[\frac{k}{R(C_{i,j}^n)h} \left(\frac{D_x}{h} - \frac{V_x}{2} \right) \right], \qquad \mathbf{B} = \left[1 - \frac{2kD_x}{R(C_{i,j}^n)h^2} - \frac{2kD_y}{R(C_{i,j}^n)h^2} \right],$$

$$\mathbf{C} = \left[\frac{k}{R(C_{i,j}^n)h} \left(\frac{D_x}{h} + \frac{V_x}{2} \right) \right], \qquad \mathbf{D} = \left[\frac{kD_y}{R(C_{i,j}^n)h} \left(\frac{1}{h} + \frac{1}{2y_i} \right) \right],$$

$$\mathbf{E} = \left[\frac{kD_y}{R(C_{i,j}^n)h} \left(\frac{1}{h} - \frac{1}{2y_i} \right) \right]. \qquad (5.13)$$

Equation (5.11) can also be expressed as:

$$C_{i,j}^{n+1} = AC_{i+1,j}^{n} + BC_{i,j}^{n} + CC_{i-1,j}^{n} + DC_{i,j+1}^{n} + EC_{i,j-1}^{n} . {(5.14)}$$

This is a finite difference scheme to equation (5.0).

CHAPTER 6

6.0 METHODOLOGY OF CALCULATING TRUNCATION ERRORS.

Writing $C_{i,j}^n$ for the value of exact solution and C(x,y,t) as the true value, we know that $C_{i,j}^n$ and $c_{i,j}^n$ satisfy the respective equation

$$R(C_{i,j}^{n})h[C_{i,j}^{n+1} - C_{i,j}^{n}] = k\left[\frac{D_{x}}{h} - \frac{V_{x}}{2}\right]C_{i+1,j}^{n} - \frac{2k}{h}\left[D_{x} + D_{y}\right]C_{i,j}^{n} + k\left[\frac{D_{x}}{h} + \frac{V_{x}}{2}\right]C_{i-1,j}^{n} + kD_{y}\left[\frac{1}{h} + \frac{1}{2y_{j}}\right]C_{i,j+1}^{n} + kD_{y}\left[\frac{1}{h} - \frac{1}{2y_{j}}\right]C_{i,j-1}^{n},$$

$$(6.0)$$

$$R(c_{i,j}^{n})h[c_{i,j}^{n+1} - c_{i,j}^{n}] = k\left[\frac{D_{x}}{h} - \frac{V_{x}}{2}\right]c_{i+1,j}^{n} - \frac{2k}{h}c_{i,j}^{n} + k\left[\frac{D_{x}}{h} + \frac{V_{x}}{2}\right]c_{i-1,j}^{n} + kD_{y}\left[\frac{1}{h} + \frac{1}{2y_{j}}\right]c_{i,j+1}^{n} + kD_{y}\left[\frac{1}{h} - \frac{1}{2y_{j}}\right]c_{i,j-1}^{n} + \Delta tT_{i,j}^{n} ,$$

$$(6.1)$$

where $T_{i,j}^n$ is the truncation error.

It is not practical to simply subtract these equations to obtain $e_{i,j}^n$, given the coefficients of R (•) are different. We can first write.

$$R(c_{i,j}^n) = R(C_{i,j}^n) + (c_{i,j}^n - C_{i,j}^n)$$

$$\tag{6.2}$$

$$=R\left(C_{i,j}^{n}\right)-e_{i,j}^{n}q_{i,j}^{n}\tag{6.3}$$

Where,
$$q_{i,j}^n = \frac{\partial R}{\partial C}(\eta)$$
 (6.4)

And η is some number between $C_{i,j}^n$ and $c_{i,j}^n$

We can now subtract equation (6.0) from (6.1) and obtain

$$R(C_{i,j}^{n})h(e_{i,j}^{n+1} - e_{i,j}^{n}) + he_{i,j}^{n}q_{i,j}^{n}\left[c_{i,j}^{n+1} - C_{i,j}^{n}\right] = k\left[\frac{D_{x}}{h} - \frac{V_{x}}{2}\right]e_{i,j}^{n} - \frac{2k}{h}\left[D_{x} + D_{y}\right]e_{i,j}^{n} + k\left[\frac{D_{x}}{h} + \frac{V_{x}}{2}\right]e_{i-1,j}^{n} + kD_{y}\left[\frac{1}{h} + \frac{1}{2y_{j}}\right]e_{i,j+1}^{n} + kD_{y}\left[\frac{1}{h} - \frac{1}{2y_{j}}\right]e_{i,j-1}^{n} - \Delta tT_{i,j}^{n}, (6.5)$$

$$e_{i,j}^{n+1} = \frac{k}{R(C_{i,j}^n)h} \left[\frac{D_x}{h} - \frac{D_x}{2} \right] e_{i+1,j}^n + \left[1 - \frac{2k}{R(C_{i,j}^n)h^2} \left(D_x + D_y \right) \right] e_{i,j}^n + \frac{k}{R(C_{i,j}^n)h} \left[\frac{D_x}{h} + \frac{V_x}{2} \right] e_{i-1,j}^n$$

$$+\frac{kD_{y}}{R(C_{i,j}^{n})h}\left[\frac{1}{h}+\frac{1}{2y_{j}}\right]e_{i,j+1}^{n}+\frac{kD_{y}}{R(C_{i,j}^{n})h}\left[\frac{1}{h}-\frac{1}{2y_{j}}\right]e_{i,j-1}^{n}-\frac{e_{i,j}^{n}q_{i,j}^{n}}{R(C_{i,j}^{n})}(c_{i,j}^{n}-C_{i,j}^{n})-\Delta tT.$$
(6.6)

The coefficient of $e_{i+1,j}^n$, $e_{i,j}^n$, $e_{i-1,j}^n$, $e_{i,j+1}^n$, $e_{i,j-1}^n$ arising from the three terms are non negative provided

$$1 - \frac{2k}{Max.R(C_{i,j}^n)h^2} \left[D_x + D_y \right] \ge 0,$$
and
$$2k \left[D_x + D_y \right] \le h^2 Max.R(C_{i,j}^n) . \tag{6.7}$$

This is our new stability condition, and condition for the approximate to satisfy a maximum principle.

Generally it will need to be checked (and Δt adjusted) at each time step.

However, assuming that we can use a constant step Δt which satisfy for all i, j and n and that we have bounds

$$\left| C_{i,j}^{n+1} - C_{i,j}^{n} \right| \le M_t \Delta t \left| q_{i,j}^n \right| \le k,$$
 (6.8)

we can write,

$$E^{n+1} \le \left[1 + kM_t \left(\Delta x\right)^2\right] E^n + \Delta t T. \tag{6.9}$$

In our previous notation:

$$\left(1 + kM_t \left(\Delta x\right)^2\right)^n \le e^{KM_t n(\Delta x)^2} < e^{KM_t x^2} \tag{6.91}$$

This allows a global error bound obtained in terms of T.

To support the model equations, our data extracted from the study carried out on soils in U.S.A i.e. Webster silty clay loam (molisol) from Iowa, Cecil sandy loam (ultisol) from Georgia, and Eutis fine sand (Entisol) from Florida. These soils were selected on the basis of their taxonomic and textural representation of major U.S.A soils. Surface samples taken from depth range of 0-30 cm depth of each soil were dried and passed through a 2 mm sieve prior to storage and use. The information taken from the detailed account is the relevant to our mathematical model and not to give irrelevant information that will make our work to become amorphous. Selected physical and chemical properties of these soils pertinent to this study are listed.

TABLE 1: Physical and chemical properties of soil used in this study

Particle size fraction (%)

Soil	Sand	Silt	Clay	PH (1.1 paste)
Webster	18.4	45.3	7.3	6.5
Cecil	65.8	19.5	14.7	4.8
Eustis	93.8	3.0	3.2	4.1

We can use the locally available soil samples if the relevant properties have been experimentally determined Pesticides:

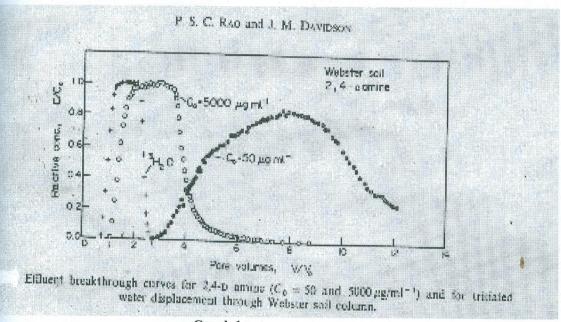
Four pesticides used in this study were 2, 4-d [2,4 Dichlorophenoxyacetic acid], atrazine [2- chloro- 4 – ethylamino – 6 – Isopylamino – 5 – triazine], terbacil [3-tert – butly – 5 – chloro – 6 – Methyluracil], and methyl parathion [0 – 0 – dimethly – 0 – p – nitrophenly phosphorothioate]

Column Displacement experiments (Relevant information)

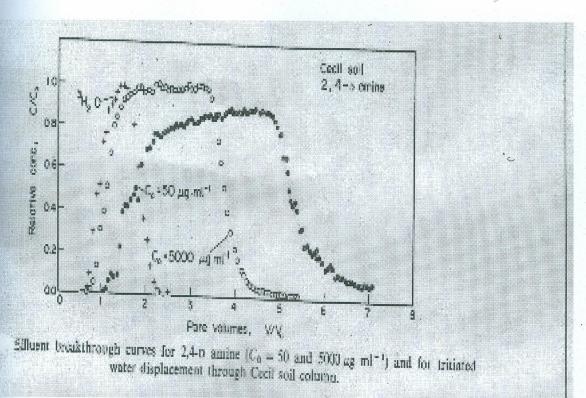
- Pesticides movement through saturated columns of Webster, cecil and Eutis soils was studied using miscible displacement technique [Davidson et al 1968 [18]
- Air dried soils were packed in small increments into glass cylinders (15 cm long: 45 cm squared cross sectional area)
- Medium porosity fitted glass end plates served to retain the soil in column.
- A known volume of pesticide solution at a desired concentration was introduced into soil at a constant flux using a constant volume peristaltic pump.
- The column experiments consisted of displacing 2, 4 damine solution at two concentrations (i.e. 50 and
- 500 $\mu g M l^{-1}$) through the columns of cecil, Eutis and webster's soil and 5 to 50 $\mu g M l^{-1}$ of atrazine through Eustis soil.
- All displacements were performed at a Darcy flux of approximately 0.22 cm/h to ensure equilibrium condition of pesticide adsorption during flow.
- The volume of water held in the soil column V_0 was gravimetrically determined at the end of each displacement by extruding the soil from glass cylinders and over drying

• The number of pore volume $\left(\frac{V}{V_0}\right)$ was calculated by dividing the cumulative outflow volume (V) by the total water volume V_0 I the soil column. Effluent pesticide concentration is expressed at relative effluent and input concentration $\left(\frac{C}{C_0}\right)$ where C and C₀ are, relatively. Plots of $\frac{C}{C_0}$ vs $\frac{V}{V_0}$ referred to as break through curves (or BTC)

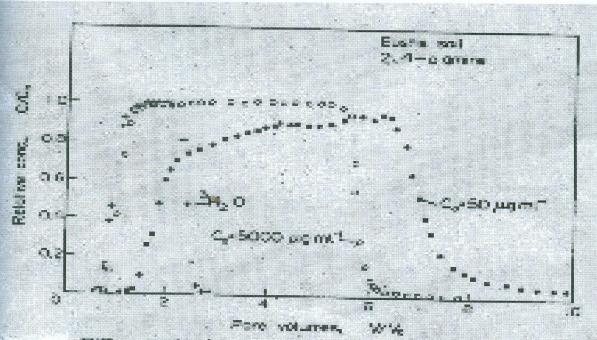
BTC. 4 Numbers Graphs



Graph 1

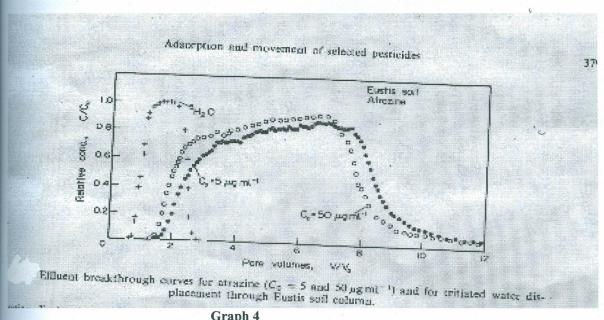


Graph 2



Effluent breakthrough curves for 2.4-p amine $(C_0 = 50 \text{ and } 5000 \, \mu \text{g m})^{-1}$ and for tritiated water displacement through Eastis soil column.

Graph 3



Graph 4

Shows Freundlich constants calculated from equilibrium adsorption isotherm for various soil TABLE 2: pesticide combination.

Pesticide	Soil	Kd	N
2, 4 – d amine	Webster	4.62	0.70
	Cecil	0.65	0.83
	Eustis	0.76	0.76
Atrazine	Webster	6.03	0.73
	Cecil	0.89	1.04
	Eustis	0.62	0.79
Terbacil	Webster	2.46	0.88
	Cecil	0.38	0.99
	Eustis	0.12	0.88
Methly Parathion	Webster	13.39	0.75
	Cecil	3.95	0.85
	Eustis	2.72	0.86

In this analysis we will use 2, 4 - D amine on webster soil to quality our mathematical model.

Data extracted and calculated.

Generally at low rates of flow the effect of molecular diffusion predominate and cell mixing contributes relatively

to dispersion. But in liquids, molecular diffusion is insignificant at Reynold number upto unity [31]

Whatever the mechanism, however, the rate of dispersion can conveniently be described by dispersion coefficient. The dispersion rate in longitudinal and radial direction is represented by D_L and D_R respectively used in representing the behavior in two directions. The process is normally linear, with rates of dispersion proportional to the products of the corresponding coefficients and concentration gradients.

Reynold number
$$=\frac{U_{c}d\rho}{\mu}$$
,
Peclet number $=\frac{Uc}{e}\frac{d}{D_{Lor}D_{R}}$,
Schimidt number $=\frac{\mu}{\rho D_{Lor}D_{R}}$.

With liquids, Schimdt number is variable and is generally about three orders of magnitude greater than gases [31] From the results in the literature pertaining longitudinal dispersion in liquids [31], it is shown that over a range of Reynolds number studied $(10^2 < \text{Rec} < 10^3)$ Peclect number show little variation and is of order of unity. Given the sieve size used in experiment is of size 2 mm, our particle diameter is 2 mm. Void ration for such size of packing is =0.402.

Taking our cylindrical vessel to be full,

Total volume =
$$15 \times 45 = 675 \text{ cm}^3$$

$$\frac{V_V}{V_S} = 0.402$$
 where V_V – volume of voids
$$V_S$$
 – volume of solids
$$V_T = V_{V+} V_S$$

$$\frac{V_T - V_S}{V_S} = 0.402$$

$$V_T = (1.402)V_S$$

$$V_S = 482cm^3$$

• Given that the soil is saturated 100% volume of water in the soil =675 - 482

$$=193 \text{ cm}^3$$

• Darcy's flux = 0.22 cm/hr,

$$V = \frac{Q}{\theta} = \frac{0.22}{0.402} = 0.547 cm/hr$$

where V is the pore water velocity.

• From our earlier literature, peclet number in liquids is approximately equals to unity despite the varying Reynolds no

$$Pe = \frac{Ucd}{eDL} = \frac{0.547(0.2)}{DL}$$

$$D_L = 0.547x0.2 = 0.11cm^2 / hr$$

Based on the same concept

$$D_R = 0.11cm^2 / hr$$

• From the break through curve [18] Webster soil; results with 2-4-Δ amine pesticide

$$\frac{V}{V_0} = 1.05$$
 for 5000 μgml^{-1}

$$\frac{V}{V_0} = 2.75$$
 for $50 \, \mu gml^{-1}$

where V is the amount of solution gone through the cylinder without pesticides i.e there was no detectable amount of pesticide in this solution meaning that the pesticide had been adsorbed completely.

Using V we can determine the amount of time taken for the adsorption to take place leaving no traces of pesticides.

$$V_0 = 193 \text{ cm}^3$$

• For pesticide concentration of 5000 μgml^{-1} ,

$$V = 1.05 \times 193 = 203 \text{ cm}^3$$

$$t = \frac{203}{45x0.22} = 20.50 \text{ hrs}$$

$$X$$
 - Section area = 45 cm²

$$Q = 0.22 \text{ cm/hr}.$$

• For
$$50 \mu gml^{-1}$$

 $V = 2.75 \times 193 = 530.75 \text{ cm}^3$
 $t = 53.6 \text{ hrs.}$

where t is the time taken for adsorption process to go on through the cylinder

Without any concentration going beyond the porous end.

$$X = 15 \text{cm}$$

 $y = 15 \text{cm}$
 $D_x = D_y = h = 0.3 \text{ cm}$
 $k = \Delta t_{(5000 \mu gml^{-1})} = 0.41 hrs$

$$k = \Delta t_{(50 \, \mu gml^{-1})} = 1.072 hrs$$

$$\rho_{_{b}} = \frac{V_{S}G + V_{0}\gamma_{W}}{V_{T}}$$

 $\gamma_{\mathcal{W}}$ - specific gravity of water.

G - specific gravity of soil = 2.68

$$=\frac{482x2.68+193x1}{675}$$

$$= 2.2 \text{ gcm}^3$$

From equation A, B, C, D and E we can determine the coefficients and subsequently $C_{i,j}^n$

$$R\left(C_{i,j}^{n}\right) = \left[1 + \frac{P_b}{\theta} KN\left(C_{i,j}^{n}\right)^{N-1}\right],$$

$$= \left[1 + 17.7\left(C_{i,j}^{n}\right) - 0.3\right].$$

$$N = 0.7$$

$$\theta = 0.402$$

$$K = 4.62$$

For $5000 \mu gml^{-1}$ pesticide coefficients,

$$A = \frac{0.12733}{R(C_{i,j}^n)} ,$$

$$\mathbf{B} = 1 - \frac{2.0004}{R(C_{i,j}^n)},$$

$$C = \frac{0.8749}{R(C_{i,j}^n)} \quad ,$$

$$E = \frac{0.15033}{R(C_{i,j}^n)} \left(\frac{1}{h} - \frac{0.5}{y_j} \right),$$

$$D = \frac{0.15033}{R(C_{i,j}^{n})} \left(\frac{1}{h} + \frac{0.5}{y_{j}}\right),$$

$$\Delta t = k = 0.41 hrs$$

$$\Delta x = \Delta y = h = 0.30cm$$

For $50\mu gml^{-1}$ pesticide concentration

$$A = \frac{0.33292}{R\left(C_{i,j}^{n}\right)} \quad ,$$

$$C = \frac{2.2875}{R(C_{i,j}^n)},$$

$$E = \frac{0.39307}{R(C_{i,j}^n)} \left(\frac{1}{h} - \frac{0.5}{y_j} \right).$$

$$B = 1 - \frac{5.2489}{R\left(\binom{n}{i,j}\right)},$$

$$D = \frac{0.39307}{R(C_{i,j}^{n})} \left(\frac{1}{h} + \frac{0.5}{y_{j}}\right),$$

The general formula for determining concentration with time is

$$C_{I,J}^{n+1} = AC_{i+1,j}^{n}BC_{i,j}^{n} + CC_{i-1,j}^{n} + DC_{i,j+1}^{n} + EC_{i,j-1}^{n}.$$

$$\Delta t = k = 1.072 hrs$$

$$\Delta x = \Delta y = h = 0.30cm$$

m the equation 5.0, we get the coefficients for one dimension equation analysis.

$$A = \frac{k}{R(C_I^n)} \left(\frac{D}{h} - \frac{V}{2} \right),\,$$

$$B = \left(1 - \frac{2kD}{R(C_i^n)h^2}\right),\,$$

$$C = \frac{k}{R(C_i^n)h} \left(\frac{D}{h} + \frac{V}{2}\right),\,$$

$$C_i^{n+1} = AC_{i+1}^n + BC_i^n + CC_{i-1}^n.$$

			ONE D	IMENSION	AL ANALYSIS $\mu gml^{-1}(C_i^n)$	$\binom{n}{i}$ $ R(C_i^n)$	
					μgπι ι		
Sno.	X	t/hrs	h/cm	Dt/k/hrs	Pesticide Concentration		
1	0.00	0	0	0	50		
2	0.30	1.072	0.3	1.072	49	6.5070	
3	0.60	2.144	0.3	1.072	48	6.5412	
4	0.90	3.216	0.3	1.072	47	6.5412	
5	1.20	4.288	0.3	1.072	46	6.5763	
6	1.50	5.36	0.3	1.072	45	6.6124	
7	1.80	6.432	0.3	1.072	44	6.6495	
8	2.10	7.504	0.3	1.072	43	6.6877	
9	2.40	8.576	0.3	1.072	42	6.7271	
10	2.70	9.648	0.3	1.072	41	6.7676	
11	3.00	10.72	0.3	1.072	40	6.8095	
12	3.30	11.792	0.3	1.072	39	6.8527	
13	3.60	12.864	0.3	1.072	38	6.8973	
	1						

	12.0		5 5			
14	3.90	13.936	0.3	1.072	37	6.9434
15	4.20	15.008	0.3	1.072	36	6.9912
16	4.50	16.08	0.3	1.072	35	7.0406
17	4.80	17.152	0.3	1.072	34	7.0919
18	5.10	18.224	0.3	1.072	33	7.1451
19	5.40	19.296	0.3	1.072	32	7.2004
20	5.70	20.368	0.3	1.072	31	7.2579
21	6.00	21.44	0.3	1.072	30	7.3178
22	6.30	22.512	0.3	1.072	29	7.3802
23	6.60	23.584	0.3	1.072	28	7.4455
24	6.90	24.656	0.3	1.072	27	7.5137
25	7.20	25.728	0.3	1.072	26	7.5851
26	7.50	26.8	0.3	1.072	25	7.6601
27	7.80	27.872	0.3	1.072	24	7.7389
28	8.10	28.944	0.3	1.072	23	7.8220
29	8.40	30.016	0.3	1.072	22	7.9096
30	8.70	31.088	0.3	1.072	21	8.0080
31	9.00	32.16	0.3	1.072	20	8.2055
32	9.30	33.232	0.3	1.072	19	8.3172
33	9.60	34.304	0.3	1.072	18	8.4369
34	9.90	35.376	0.3	1.072	17	8.5655
35	10.20	36.448	0.3	1.072	16	8.7044
36	10.50	37.52	0.3	1.072	15	8.8550
37	10.80	38.592	0.3	1.072	14	9.0193
38	11.10	39.664	0.3	1.072	13	9.1996
39	11.40	40.736	0.3	1.072	12	9.3988
40	11.70	41.808	0.3	1.072	11	9.6210
41	12.00	42.88	0.3	1.072	10	9.8710
42	12.30	43.952	0.3	1.072	9	10.1589
43	12.60	45.024	0.3	1.072	8	10.4852
44	12.90	46.096	0.3	1.072	7	10.8729
45	13.20	47.168	0.3	1.072	6	11.3402
46	13.50	48.24	0.3	1.072	5	11.9215
47	13.80	49.312	0.3	1.072	4	12.6777
48	14.10	50.384	0.3	1.072	3	13.7363
50	14.70	52.528	0.3	1.072	1	18.7000
51	15.00	53.6	0.3	1.072	0	1.0000
With the						

CONCI	CONCENTRATION AND COEFFICIENT OF ONE DIMENSIONAL ANALYSIS									
Sno.	A	В	С		C_i^{n+1}					
1	0	0	0							
2	0.05112	0.597295	0.35155		49.300					
3	0.05089	0.5994	0.34971		48.299					
4	0.05062	0.60154	0.34784		47.297					
5	0.05035	0.60371	0.34594		46.296					
6	0.05006	0.60593	0.3440		45.294					
7	0.04978	0.6082	0.34200		44.291					
8	0.04949	0.61047	0.3400		43.291					
9	0.04919	0.612802	0.3380		42.288					
10	0.04889	0.615185	0.33603		41.291					
11	0.04857	0.61761	0.33381		40.284					
12	0.04827	0.62008	0.33282		39.330					
13	0.04795	0.62261	0.32945		38.282					
14	0.04976	0.6252	0.32720		37.357					
15	0.04727	0.62782	0.32490		36.277					
16	0.04694	0.63051	0.32255		35.275					
17	0.04659	0.63326	0.3202		34.274					
18	0.046233	0.63608	0.31769		33.272					
19	0.045867	0.638959	0.31709		32.269					
20	0.045492	0.64191	0.31259		31.267					
21	0.045432	0.644966	0.30995		30.266					
22	0.04311	0.64806	0.30723		29.262					
23	0.04471	0.65125	0.30444		28.260					
24	0.043889	0.654533	0.30158		27.252					
25	0.043459	0.657916	0.29862		26.255					
26	0.043016	0.6614	0.29558		25.253					
27	0.042560	0.664996	0.292444		24.248					
28	0.042088	0.66871	0.28921		23.247					
29	0.042000	0.67255	0.28585		22.244					
30	0.041095	0.67653	0.28238		21.241					
31	0.041093	0.68065	0.27878		20.238					
32	0.04037	0.68494			19.235					
33	0.039458	0.68947	0.27503 0.27113							
34	0.038865	0.69408	0.26706		18.233 17.228					
35	0.038245	0.69896	0.26279		16.225					
36	0.03759	0.70408	0.25833		15.221					
37	0.03739	0.70947	0.25362		14.217					
38	0.036186	0.71516	0.24865		13.213					
39	0.03542	0.7212	0.24338							
40	0.03460	0.72763	0.23776		12.208 11.203					
41	0.03373	0.72703	0.23174		10.198					
42	0.03277	0.74206	0.22517		9.192					
43	0.03277	0.75005	0.21817							
44	0.03175	0.75005	0.21017		8.187					
45	0.030616	0.75899	0.21039		7.180					
46	0.029336	0.7802	0.20172		6.172					
47	0.62626	0.7802			5.164					
48	0.02626		0.18043		4.154					
49	0.02425	0.80915	0.16660		3.142					
50	0.021649	0.82959 0.85987	0.14876		2.127					
33	0.017002	0.00301	0.12233		1.105					
				30						

51 | 0.000

ONE DIMENSIONAL ANALYSIS $\mu gml^{-1}(C_i^n)$								
Sno.	×	t/hrs	h/cm	Dt/k/hrs	Pesticide Concentration	$R(C_i^n)$		
1	0.00	0	0	0	5000			
2	0.30	0.41	0.3	0.41	4900	2.3833		
3	0.60	0.82	0.3	0.41	4800	2.3919		
4	0.90	1.23	0.3	0.41	4700	2.4007		
5	1.20	1.64	0.3	0.41	4600	2.4098		
6	1.50	2.05	0.3	0.41	4500	2.4191		
7	1.80	2.46	0.3	0.41	4400	2.4287		
8	2.10	2.87	0.3	0.41	4300	2.4386		
9	2.40	3.28	0.3	0.41	4200	2.4488		
10	2.70	3.68	0.3	0.41	4100	2.4593		
11	3.00	4.10	0.3	0.41	4000	2.4701		
12	3.30	4.51	0.3	0.41	3900	2.4813		
13	3.60	4.92	0.3	0.41	3800	2.4929		
14	3.90	5.33	0.3	0.41	3700	2.5049		
15	4.20	5.74	0.3	0.41	3600	2.5173		
16	4.50	6.15	0.3	0.41	3500	2.5302		
17	4.80	6.56	0.3	0.41	3400	2.5436		
18	5.10	6.97	0.3	0.41	3300	2.5575		
19	5.40	7.38	0.3	0.41	3200	2.5719		
20	5.70	7.79	0.3	0.41	3100	2.5869		
21	6.00	8.2	0.3	0.41	3000	2.6026		
22	6.30	8.61	0.3	0.41	2900	2.6190		
23	6.60	9.02	0.3	0.41	2800	2.6362		
24	6.90	9.43	0.3	0.41	2700	2.6541		
25	7.20	9.84	0.3	0.41	2600	2.6729		
26	7.50	10.25	0.3	0.41	2500	2.6927		
27	7.80	10.66	0.3	0.41	2400	2.7136		
28	8.10	11.07	0.3	0.41	2300	2.7353		
29	8.40	11.48	0.3	0.41	2200	2.7589		
30	8.70	11.89	0.3	0.41	2100	2.7836		
31	9.00	12.30	0.3	0.41	2000	2.8099		
32	9.30	12.71	0.3	0.41	1900	2.838		
33	9.60	13.12	0.3	0.41	1800	2.8681		
34	9.90	13.53	0.3	0.41	1700	2.9004		
35	10.20	13.94	0.3	0.41	1600	2.9353		
36	10.50	14.35	0.3	0.41	1500	2.9731		
37	10.80	14.76	0.3	0.41	1400	3.014		
38	11.10	15.17	0.3	0.41	1300	3.0596		
39	11.40	15.58	0.3	0.41	1200	3.1097		
40	11.70	15.99	0.3	0.41	1100	3.1655		
41	12.00	16.40	0.3	0.41	1000	3.2283		
42	12.30	16.81	0.3	0.41	900	3.2999		
43	12.60	17.22	0.3	0.41	800	3.3826		
			7.0	7		0.0020		

44	12.90	17.63	0.3	0.41	700	3.48
45	13.20	18.04	0.3	0.41	600	3.5973
46	13.50	18.45	0.3	0.41	500	3.7434
47	13.80	18.86	0.3	0.41	400	3.9333
48	14.10	19.27	0.3	0.41	300	4.1977
49	14.10	19.68	0.3	0.41	200	4.6113
50	14.70	20.09	0.3	0.41	100	5.446
51	15.00	20.50	0.3	0.41	0	1.0000

CONCENTRATION AND COEFFICIENT OF ONE DIMENSIONAL ANALYSIS

	A			$-C_i^{n+1}$
-	Α			
		В	C	
1 1				
2	0.053413	0.57949	0.36710	4931.38
	0.05322	0.58100	0.36578	4831.26
	0.053026	0.58254	0.36444	4731.17
	0.052826	0.58411	0.363059	4631.00
	0.052623	0.58571	0.363059	4530.89
	0.05242	0.58735	0.36023	4430.76
The state of the s	0.052202	0.58903	0.35877	4330.67
The second secon	0.051985	0.59074	0.357277	4230.54
10	0.051763	0.592486	0.355772	4130.40
11	0.051564	0.594267	0.354196	4030.37
12	0.051304	0.596099	0.352597	3930.13
13	0.051065	0.597978	0.350957	3829.99
14	0.050820	0.599904	0.349275	3729.84
15	0.0505701	0.601875	0.347555	3629.70
	0.050312	0.603905	0.34578	3529.54
17	0.050047	0.60599	0.34396	3429.81
18	0.049775	0.608133	0.34209	3329.23
19	0.49497	0.610327	0.34018	3229.08
20	0.049209	0.612587	0.33820	3129.76
21	0.049209	0.614924	0.336164	3028.73
22	0.048606	0.617335	0.334055	2928.55
23	0.04829	61983	0.33188	2828.36
24	0.047964	0.62052	0.331276	2727.68
25	0.04763	0.625051	0.32732	2627.97
26	0.04728	0.627809	0.32492	2527.79
27	0.046912	0.63068	0.32241	2427.56
28	0.04654	0.63361	0.319855	2427.34
29	0.04614	0.63674	0.31712	2227.10
30	0.04573	0.63996	0.3143	2126.85
31	0.0453	0.64333	0.31136	2026.59
32	0.04486	0.64686	0.3083	1926.34
33	0.04438	0.6506	0.30505	1826.11
34	0.04389	0.65446	0.30165	1725.77
35	0.04337	0.65856	0.29806	1625.46
36	0.04282	0.66291	0.29427	1525.15
37	0.04224	0.66749	0.29028	1424.82
38	0.041607	0.67244	0.28595	1324.43

0.040936	0.67772	0.28135		1224.04
0.040215	0.683399	0.276386		1123.62
0.03943	0.68956	0.27101		1023.16
0.038577	0.69629	0.26513		922.65
0.03763	0.70372	0.25865		822.10
0.03658	0.7120	0.25141		721.47
0.03539	0.7214	0.24321		620.78
0.034006	0.7322	0.23372		519.93
0.030326	0.7452	0.22243		419.00
0.030326	0.76125	0.20842		317.81
0.027606	0.78266	0.18973	0	216.21
0.23375	0.815975	0.16065	B R	113.72
				0.00

	ANALY	MENSIONAL SIS	T. Pr	8	₁ -1n.	$n \in \mathbb{R}^n$
	7 × - 11			$k = \Delta t$	$\mu gml^{-1}(C^ni,$	(C^n)
Cno.	$x_i = y_j$	4/lowe	la /a ma		Destinide concentration	
Sno.		t/hrs	h/cm		Pesticide concentration	
1	0	0	0	0	50	4.0 2.0.0
2	0.30	1.072	0.3	1.072	49	6.507
3	0.60	2.144	0.3	1.072	48	6.5412
1	0.90	3.216	0.3	1.072	47	6.5763
5	1.20	4.288	0.3	1.072	46	6.6124
3	1.50	5.360	0.3	1.072	45	6.6495
7	1.80	6.432	0.3	1.072	44	6.6877
3	2.10	7.504	0.3	1.072	43	6.7271
9	2.40	8.576	0.3	1.072	42	6.7676
10	2.70	9.648	0.3	1.072	41	6.8095
11	3.00	10.720	0.3	1.072	40	6.8527
2	3.30	11.792	0.3	1.072	39	6.8973
3	3.60	12.864	0.3	1.072	38	6.9434
14	3.90	13.936	0.3	1.072	37	6.9912
5	4.20	15.008	0.3	1.072	36	
	1					7.0406
6	4.50	16.080	0.3	1.072	35	7.0919
7	4.80	17.152	0.3	1.072	34	7.1451
8	5.10	18.224	0.3	1.072	33	7.2004
19	5.40	19.296	0.3	1.072	32	7.2579
20	5.70	20.368	0.3	1.072	31	7.3178
21	6.00	21.440	0.3	1.072	30	7.3802
22	6.30	22.512	0.3	1.072	29	7.4455
23	6.60	23.584	0.3	1.072	28	7.5137
24	6.90	24.656	0.3	1.072	27	7.5851
25	7.20	25.728	0.3	1.072	26	7.6601
26	7.50	26.800	0.3	1.072	25	7.7389
27	7.80	27.872	0.3	1.072	24	7.8220
28	8.10	28.944	0.3	1.072	23	7.9096
29	8.40	30.016	0.3	1.072	22	8.0024
80	8.70	31.088	0.3	1.072	21	8.1008
31	9.00	32.160	0.3	1.072	20	8.2055
32	9.30	33.232	0.3	1.072	19	8.3172
33	9.60	34.304	0.3	1.072	18	8.4369
34	9.90	35.376	0.3	1.072	17	
35	10.20	36.448	0.3		1000	8.5655
	The second secon	1 2		1.072	16	8.7044
6	10.50	37.52	0.3	1.072	15	8.8550
37	10.80	38.592	0.3	1.072	14	9.0193
38	11.10	39.664	0.3	1.072	13	9.1996
39	11.40	40.736	0.3	1.072	12	9.3988
10	11.70	41.808	0.3	1.072	11	9.6210
11	12.00	42.88	0.3	1.072	10	9.8710
12	12.30	43.952	0.3	1.072	9	10.1589
13	12.60	45 024	0.2	1 072		40 4050

10.4852

10.8729

11.3402

11.9215

12.6777

8

7

6

5

4

1.072

1.072

1.072

1.072

1.072

43

44

45

46

47

12.60

12.90

13.20

13.50

13.80

45.024

46.096

47.168

48.24

49.312

0.3

0.3

0.3

0.3

0.3

			200			
48	14.10	50.384	0.3	1.072	3	13.7303
49	14.10	51.456	0.3	1.072	2	15.3769
50	14.70	52.528	0.3	1.072	. 1	18.7000
51	15.00	53.60	0.3	1.072	0	1.0000

$$2(1.072)(0.22) \le 0.3^2(18.7)$$

 $0.47168 \le 1.683$

CONC	CONCENTRATION AND COEFFICIENT OF TWO DIMENSIONAL ANALY C n+1									
Sno.	٨	В	С	D	E	$[i, j\mu g / ml]$				
1	A	В	C	D	<u> </u>	50				
2	0.051163	0.19335	0.35155	0.30204	0.10068	49.04				
3	0.051103	0.19355	0.33133	0.30204	0.15023	48.14				
4	0.05090	0.197303	0.34794	0.23244	0.15023	47.21				
5	0.05002	0.20103	0.34594	0.23244	0.10003	46.19				
6	0.05007	0.2002	0.34401	0.22292	0.17336	45.20				
7	0.03007	0.215141	0.342046	0.21073	0.17734	44.06				
8	0.04949	0.21974	0.342040	0.210961	0.17981	43.22				
9	0.04949	0.21974	0.34004	0.205704	0.180657	42.22				
10	0.04919	0.22918	0.33593	0.203704	0.181723					
11	0.048582	0.22910	0.33381	0.20310		41.20				
12	0.048362	0.23417	0.33361	0.20076	0.18164	40.22				
13	0.046266	0.236992	0.32945	0.195555	0.181329	39.22				
14	0.047946	0.249213	0.32945	0.195555	0.180840	38.18				
15	0.04702	0.249213		The second of th	0.180204	37.20				
16	0.04729	0.254461	0.32490 0.32255	0.192743	0.179451	36.22				
	0.046594			0.1909091	0.17859	35.22				
17		0.26539	0.32015	0.18911	0.17765	34.23				
18	0.046236	0.27103	0.31770	0.18732	0.17661	33.22				
19	0.045870	0.2768	0.315179	0.18554	0.17551	32.22				
20	0.045495	0.28272	0.312595	0.183597	0.17434	31.22				
21	0.04511	0.28879	0.309951	0.181972	0.1731	30.22				
22	0.044714	0.295024	0.30723	0.180167	0.171787	29.22				
23	0.04431	0.301423	0.304444	0.178342	0.170416	28.22				
24	0.0438913	0.307999	0.301578	0.17493	0.168983	27.18				
25	0.043462	0.314774	0.298625	0.17461	0.167483	26.26				
26	0.043019	0.321751	0.295585	0.172691	0.1659187	25.22				
27	0.042562	0.32956	0.292444	0.170727	0.1642849	24.22				
28	0.042091	0.3363887	0.289206	0.168719	0.16251	23.21				
29	0.0416025	0.3340843	0.2858517	0.166654	0.1608063	22.22				
30	0.04109717	0.352052	0.2823795	0.16453	0.1589526	21.21				
31	0.0405728	0.360319	0.278776	0.1623387	0.156758	20.21				
32	0.0400279	0.36891	0.27503	0.160074	0.15499	19.21				
33	0.03946	0.377864	0.27113	0.157725	0.15287	18.21				
34	0.038868	0.038868	0.26706	0.15528	0.1506487	17.21				
35	0.038247	0.396983	0.262798	0.152739	0.148312	16.21				
36	0.037597	0.40724	0.25833	0.15008	0.145852	15.20				
37	0.036912	0.418037	0.2536623	0.147288	0.143252	14.20				
38	0.0361887	0.42944	0.248653	0.144348	0.140499	13.20				
39	0.036099	0.441535	0.243382	0.141239	0.13757	12.20				
40	0.034604	0.454433	0.237761	0.137931	0.134439	11.19				
41	0.0337271	0.468250	0.231739	0.134395	0.131076	10.19				
42	0.032771	0.48332	0.225172	0.130547	0.12740	9.18				
43	0.031751	0.49940	0.218165	0.126448	0.12347	8.18				
44	0.030619	0.517250	0.210385	0.121906	0.11910	7.17				
45	0.0293575	0.537142	0.201716	0.116852	0.114226	6.17				
46	0.027926	0.559711	0.191880	0.111126	0.108684	5.16				
47	0.0262603	0.585973	0.180435	0.104473	0.102226	4.15				
48	0.024247	0.617714	0.166602	0.096442	0.04411	3.14				
49	0.021651	0.658650	0.148762	0.086095	0.084320	2.12				
50	0.017803	0.71931	0.122326	0.070781	0.069351	1.10				
51		,			2.230001	0.00				
					46	3.00				

TWO DIMENSIONAL ANALYSIS								
				$k = \Delta t$	$\mu gml^{-1}(C_{i,j}^{n},j)$	$R(C^n i, j)$		
	$ \mathbf{r} - \mathbf{v} $							
Sno.	$x_i = y_j$	t/hrs	h/cm		Pesticide concentration			
1	0.00	0.00	0	0	5000			
2	0.30	0.41	0.3	0.41	4900	2.3833		
3	0.60	0.82	0.3	0.41	4800	2.3919		
4	0.90	1.23	0.3	0.41	4700	2.4007		
5	1.20	1.64	0.3	0.41	4600	2.4098		
6	1.50	2.05	0.3	0.41	4500	2.4191		
7	1.80	2.46	0.3	0.41	4400	2.4287		
8	2.10	2.87	0.3	0.41	4300	2.4386		
9	2.40	3.28	0.3	0.41	4200	2.4488		
10	2.70	3.69	0.3	0.41	4100	2.4593		
11	3.00	4.10	0.3	0.41	4000	2.4701		
12	3.30	4.51	0.3	0.41	3900	2.4813		
13	3.60	4.92	0.3	0.41	3800	2.4929		
14	3.90	5.33	0.3	0.41	3700	2.5049		
15	4.20	5.74	0.3	0.41	3600	2.5173		
16	4.50	6.15	0.3	0.41	3500	2.5302		
17	4.80	6.56	0.3	0.41	3400	2.5436		
18	5.10	6.97	0.3	0.41	3300	2.5436		
19	5.40	7.38	0.3	0.41	3200	2.5719		
20	5.70	7.79	0.3	0.41	3100			
21	6.00	8.20	0.3	0.41	3000	2.5869		
22	6.30	8.61	0.3	0.41	2900	2.6026		
23	6.60	9.02	0.3	0.41		2.6190		
24	6.90	9.02			2800	2.6362		
25	7.20		0.3	0.41	2700	2.6541		
26		9.84	0.3	0.41	2600	2.6729		
27	7.50	10.25	0.3	0.41	2500	2.6927		
	7.80	10.66	0.3	0.41	2400	2.7136		
28	8.10	11.07	0.3	0.41	2300	2.7356		
29	8.40	11.48	0.3	0.41	2200	2.7589		
30	8.70	11.89	0.3	0.41	2100	2.7836		
31	9.00	12.30	0.3	0.41	2000	2.8099		
32	9.30	12.71	0.3	0.41	1900	2.838		
33	9.60	13.12	0.3	0.41	1800	2.8681		
34	9.90	132.53	0.3	0.41	1700	2.9004		
35	10.20	13.94	0.3	0.41	1600	2.9353		
36	10.50	14.35	0.3	0.41	1500	2.9731		
37	10.80	14.76	0.3	0.41	1400	3.014		
38	11.10	15.17	0.3	0.41	1300	3.0596		
39	11.40	15.58	0.3	0.41	1200	3.1097		
40	11.70	15.99	0.3	0.41	1100	3.1655		
41	12.00	16.40	0.3	0.41	1000	3.2283		
42	12.30	16.81	0.3	0.41	900	3.2999		
43	12.60	17.22	0.3	0.41	800	3.3826		

	ř .	i .	ř .	1	1	
44	12.90	17.63	0.3	0.41	700	3.4800
45	13.20	18.04	0.3	0.41	600	3.5973
46	13.50	18.45	0.3	0.41	500	3.7434
47	13.80	18.86	0.3	0.41	400	3.9333
48	14.10	19.27	0.3	0.41	300	4.1977
49	14.10	19.68	0.3	0.41	200	4.6113
50	14.70	20.09	0.3	0.41	100	5.4460
51	15.00	20.50	0.3	0.41	0	1

$$2 k [Dx + Dy] \le h^{2} MaxR (C_{i,j}^{n})$$

$$MaxR (C_{i,j}) = 5.446$$

$$Dx = Dy = 0.11 cm^{2} / hr$$

$$2 (0.41)(0.11 + 0.11) \le (0.3)^{2} (5.446)$$

$$0.1804 \le 0.49014$$

CONCENTRATION AND COEFFICIENT OF TWO DIMENSIONAL

Sno.	Δ	В	С	D	E	$C_{i,j}^{n+1} / \mu gml^{-}$
	A	В	C	D	_	5000
1	0.0524	0.15000	0.3671	0.2454	0.1051	5000
2	0.0534	0.15898 0.16201	0.3658	0.3154	0.1051	4910
100	20.0			0.2619	0.1571	4821
4	0.0530	0.1651	0.3644	0.2435	0.17394	4748
5	0.0524	0.16823 0.1716	0.3631	0.2339	0.1819	4626
6	0.05264		0.3617	0.2279	0.1864	4528
7	0.05242	0.1747	0.3602	0.2235	0.18913	4427
3	0.05221	0.1781	0.3588	0.2202	0.1908	4328
9	0.05199	0.1815	0.3573	0.2174	0.1918	4228
10	0.05178	0.18497	0.35575	0.2151	0.1924	4128
11	0.05155	0.18853	0.354196	0.21301	0.19272	4028
12	0.05132	0.192198	0352597	0.21113	0.19277	3928
13	0.051077	0.19596	0.35096	0.209386	0.192635	3828
14	0.050832	0.199808	0.34928	0.20774	0.192354	3728
15	0.050582	0.20375	0.34755	0.20617	0.191953	3628
16	0.050324	0.20781	0.34578	0.20425	0.19145	3527
17	0.050059	0.21198	0.34396	0.20316	0.190847	3428
18	0.049786	0.216266	0.34209	0.201696	0.190171	3328
19	0.049508	0.220654	0.340176	0.200249	0.189424	3228
20	0.049221	0.225173	0.338204	0.198804	0.188609	3128
21	0.048924	0.23000	0.336164	0.197352	0.187725	3028
22	0.048618	0.2346697	0.33406	0.195888	0.186777	2927
23	0.048301	0.239663	0.331880	0.19440	0.185764	2828
24	0.047975	0.24479	0.329641	0.19291	0.18469	2727
25	0.047637	0.250103	0.32732	0.19138	0.18357	2627
26	0.047287	0.255617	0.32492	0.189818	0.182374	2527
27	0.046923	0.26135	0.32241	0.188214	0.181111	2427
28	0.0465455	0.26729	0.31982	0.186570	0.179785	2327
29	0.0461525	0.273479	0.317119	0.184874	0.178387	2226
30	0.045743	0.27993	0.314305	0.183122	0.176915	2126
31	0.0453148	0.286665	0.311363	0.181306	0.175356	2026
32	0.044866	0.293728	0.308281	0.179416	0.173720	1926
33	0.044395	0.30114	0.30505	0.17745	0.171985	1826
34	0.043901	0.30893	0.30165	0.175387	0.170152	1725
35	0.043379	0.31714	0.29806	0.173226	0.168205	1625
36	0.042827	0.32582	0.294272	0.170952	0.166137	1525
37	0.042246	0.33497	0.0.290279	0.168567	0.163948	1424
38	0.041617	0.344882	0.28595	0.165993	0.161566	1324
39	0.040946	0.35544	0.281345	0.16326	0.15902	1224
40	0.040224	0.366798	0.276386	0.1603299	0.156271	1123
41	0.039442	0.3791159	0.27101	0.157161	0.15328	1023
42	0.038586	0.392587	0.265129	0.153705	0.15000	922
43	0.037643	0.40744	0.25865	0.133703	0.13000	822
44	0.037643	0.424023	0.251408	0.1456686	0.140377	721
45	0.035396	0.424023	0.231408	0.1436666	50.10	
46	0.033396	0.442804	The second secon		0.137716	621
40 47			0.233712	0.1353496	0.132375	520
	0.032372	0.490402	0.22243	0.128784	0.126015	419
48	0.030333	0.5225	0.208424	0.1206448	0.118105	318

 50
 0.0233805
 0.63195
 0.16065
 0.092951
 0.091074
 114

 51
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0
 0

CHAPTER 8

CONCLUSION

1) Our calculations are in agreement with equation (6.7) i.e.

$$2K(D_{\mathcal{X}} + D_{\mathcal{Y}}) \le h^2 MaxR(C_{i,j}^n)$$

- 2) The lower the concentration of adsorbate the lower the adsorption which is qualified by the retardation faction which increases with lowering of the concentration i.e. it is inversely proportional to the concentration.
- 3) Explicit method used in one dimensional case produce negative coefficient thus confirming the instability of the method despite the, algebraic addition results but all implicit methods used in one or two dimension cases produce positive coefficients thus confirming the stability of implicit method.

RECOMMENDATIONS:

- 1) More studies need to be carried out on solving the two dimensional equation using analytical methods by introducing linearization factors [26]
- 2) We require a similar experimental study to be carried out on our locally available soils samples.

REFERENCÉS.

- 1. Bailey G. W & J. L.: (1970), Factors influencing adsorption, desorption and movement of pesticides in soil. Residue rev. 32: 29-92.
- 2. Blyth F.G, De Feitas M.H.; (1984), Geology for Engineering. (Seventh Edition)(English Language Book Society/Edward Arnold.)
- 3. Davidson J. M., OuL.T and Rao P.S.C.: (1976) *Behaviour of high pesticide concentration in the soil wate systems. In Residual management by land disposal*: Proceedings of the hazardous waste research symptoms. (Edited by W.H. Fuller) pp 206-212. EPA-600 19-76-015, (July 1976).
- 4. Faust S.D and Osman Aly.: (1983). Chemistry of water treatment. (Published by American Book Society)
- 5. Fowler A.C.: (1979).Mathematical models in applied Sciences. Cambridge text in applied mathematics. (ELBS).
- 6. Freundlick H, (1926). Colloid and capillarity chemistry (London Methuen and company.)
- 7. Hamaker J. W. & Thomson J. M.: (1972). adsorption of inorganic chemicals in soil environment. (Edited by C. A. I Gloring and J.W. Hamaker) Vol. 1 pp 49-143 Marcel Dekker, N.Y.
- 8. Jain M.K., I yenger S.R. K, Jain R. K.: (1984). Numerical Methods for Scientific and Engineering computation.
- 9. Manes M.: (1980).In: Activated carbon adsorption of organics from the aqueous phase Vol. I.V. Suffet and M.J. Mc Guire, Eds (Ann Arbor, MI: Ann Arbor Science Publishers Inc. 1980) P.43
- Morris, J. C. and Weber W.J., Jr: (1964). "Adsorption of Biochemically Resistant Materials from solution
 Environmental Health Series, AWTR 9 U.S Dept of Health, Education and Welfare, Washington D.C
- 11. Morton K.W. & Mayers D.F: (1989). Numerical Solution of Partial Differential Equation. (ELBS).
- 12. Rao P.S.C. and Davidson J.M:(1979). *Adsorption and movement of selected pesticide at high concentration in the soil water*. Research journal volume 13, Pergamon press Ltd 1979 Britain.
- 13. Sanborn J. R, Francis B.F & Metcalf R.L: (1977). The degradation of selected pesticides in soil; a review of published literature. EPA -600/9-77-022
- 14. Schomaker N.B: (1976). Current research on land disposal of hazardous wastes. In residue management

by land Disposal: Proc. of the hazardous waste res. Symp. (Edited by W.H. Fuller) EPA-600/9-76-015, p. 1-13 July)

- 15. Sokolnikoff, Redheffer:(1958). Mathematics of Physics and Modern Engineering 2nd Ed. Published by McGRAW-HILL KOGAKUSHA, LTD.
- 16. Van Genuchten M. Th, Davidson J.M & Wierenga, P. J. (1974). An Evaluation of Kinetic and equilibrium equation for predicting pesticide movement through porous media. Soil Sci. Soc. Am
- 17. Van Genuchten, M.Th and J.C. Parker: (1984). Boundary Conditions for displacement experiments through short laboratory column soil sci: soc. Am. J. 48:703-708.
- 18. Von Everdingen R.O & Freeze R.A.: (1971). Subsurface disposal of waste in Canada Tech. Bull No. 49 I 19 Inland water Branch, Department of Environment, Ottawa, Canada.
- 19. Weber, W.J. Jr Physicio: (1972). Chemistry process for waste quality control (New York: Wiley Interscience.)
- 20. Weil I, Morris J.C. (1949) Caulson and Richardson 4th Edition Chemical Engineering. (Ame. Chem society 71: 1664).
- 21. Wilson G.V., Selim H.M. and Dane J.H:(2005). Desert Research Institute Las Vegas, Louisiana state University, and Aubum University Southern Coop series Bulletin, SCSB# 395.
- 22. Yoke T.L. et al In(1980). Activated carbon Adsorption of organics from aqueous phase, Vol 2, M.J. Mc Guire and I.H. suffet. Eds (Ann Arbor, MI. Ann Arbor Science Publishers Inc. 1980; Page 27
- 23. Young D.M and Growell A.D:(1962) Physical adsorption of gases (London: Butterworth, 1962)
- 24. Zogorski, J.S.:(1975). The adsorption of phenols onto granular Activated Carbon from aqueous solution: PHD Thesis, Rutgers University, New Brunswick, NJ (1975)